



Acid Rain

- What is acid rain?
- Types of acid involved
- Acid rain precursors
- Transport and conversion of precursors
- pH of precipitation
- Biological effects
- Alkalinity
- Mitigation methods/technology
- Acid rain policy

Acid Deposition

Wet Deposition (acid rain, snow)

- Easy to measure; about half of acid deposition
- First measured in 1852 in England
- No widespread attention until 1961, when Swedish scientists hypothesized that declining fish populations (and, later, forest damage) might be caused by acid rain from coal-burning regions of Europe.
- By 1980, recognized as a major environmental problem in industrialized countries.

Acid Deposition

Dry deposition (acid gases and aerosols)

- Difficult to measure; importance recognized only in 1980s
- Main form of acid deposition near source
- Main cause of damage to foliage?

Droplet deposition (acid fogs)

- Generally less important than wet and dry deposition, except in certain locations
- Highest measured acidities, but total dose small

Acids and Acid Precursors

- Sulfur dioxide (SO₂) \rightarrow sulfuric acid (H₂SO₄)
 - primarily from coal burning and smelting
- Nitrogen oxides (NO, NO₂) \rightarrow nitric acid (HNO₃)
 - primarily from high-temperature combustion
- Formic and acetic acids (HCO₂H, CH₃CO₂H)
 - primarily from biomass burning, mostly in Africa and South America
- Carbonic acid ($CO_2 \rightarrow H_2CO_3$)
 - from CO₂ in atmosphere, responsible for acidity of pristine precipitation

Sources of Sulfur Dioxide (SO₂)

Anthropogenic

- 60-100 Mt(S)/y emitted into atmosphere worldwide
- US emissions: 32 Mt/y(1973)

 \rightarrow 9.5 (2006) Mt/y SO₂

- Coal-burning in power plants
- Oil-burning (vehicles)
- Industrial (smelting, etc.); responsible for intense local damage
- Biomass burning: 1-6 Mt(S)/y worldwide
 <u>Natural</u>
- volcanoes (6-20 Mt(S)/y as SO₂)

$1 Mt(S) = 2 Mt(SO_2) = 3 Mt(SO_4^{2})$

- Atomic mass of O = 16 g/mol
- Atomic mass of S = 32 g/mol
- Molecular mass of $SO_2 = 64 \text{ g/mol}$
- Formula mass $SO_4^{2-} = 96 \text{ g/mol}$

SO₂ to Acid

 In water solution, sulfur dioxide dissolves in water to form sulfurous acid, which can go on to form sulfuric acid.

 $SO_2 + H_2O \rightleftharpoons H_2SO_3$

 $H_2SO_3 + H_2O_2 \rightarrow H_2SO_4 + H_2O$

 In the gas phase, sulfur dioxide can form sulfuric acid, H₂SO₄.

 $SO_2 + OH \cdot \rightarrow HOSO_2 \cdot$ $HOSO_2 \cdot + O_2 \rightarrow HO_2 \cdot + SO_3$ $SO_3(g) + H_2O(I) \rightarrow H_2SO_4(I)$

Sulfurous and Sulfuric Acid Ionizations

Sulfuric acid and sulfurous acid form hydronium ions

 $H_2SO_4 + H_2O \rightarrow H_3O^+ + HSO_4^ HSO_4^- + H_2O \rightleftharpoons H_3O^+ + SO_4^{2-}$ $H_2SO_3 + H_2O \rightleftharpoons H_3O^+ + HSO_3^ HSO_3^- + H_2O \rightleftharpoons H_3O^+ + SO_3^{2-}$

• The pH can drop to 4.3 or lower.

Global Sulfur Budget





	Region	Coal Use 1990, (Mt)	Area (10 ⁶ km ²)	Density (t/km ²)	2003
Average density	United States	784	7.672	102	120
of coal use in the	New England	7.8	0.163	48	
	Middle Atlantic	80	0.258	308	
US, western	EN Central	204	0.632	322	
Europe and	WN Central	77	1.316	59	
Luiope, and	South Atlantic	158	0.691	229	
China	ES Central	93	0.463	200	
100 ± 100	WS Central	80	1.107	72	
$IS \approx 100 \text{ L/km}^2/\text{y}$	Mountain	79	2.215	36	
in high-use areas,	Pacific	6.6	0.826	8.0	70
>300 t/km²/y	Western Europe	400	4.375	91	
	France	24	0.547	44	
	BeNeLux	33	0.029	1160	
	Italy	21	0.301	68	
	Spain	25	0.505	49	
	Sweden	3.8	0.450	8.4	
	Turkey	31	0.780	40	
	United Kingdom	91	0.244	370	
	West Germany	110	0.249	440	150

Coal Use and SO₂ Emissions

- In 1990, density of coal use in US, western Europe, and China averaged ≈100 t/km²/y
- Sulfur content of US coal ranges from <0.5% to >7%; on average, ≈2.5% S
- Emission density: What is the emission density in g(S)/m²/y?

$$\frac{2 \text{ g S}}{m^2 \text{ y}} = \frac{100 \text{ t coal}}{\text{km}^2 \text{ y}} \left(\frac{2.5 \text{ t S}}{100 \text{ t coal}}\right) \left(\frac{10^6 \text{ g}}{1 \text{ t}}\right) \left(\frac{1 \text{ km}}{10^3 \text{ m}}\right)^2 = 2.5 \text{ g(S)/m}^2 \text{ y}$$

Emissions are 3+ times higher in high coal-use areas

Where SO₂ Emissions Deposited?

- Of SO₂ emitted in the Eastern US, ≈25% is converted into H₂SO₄, deposited in rain/snow
- The residence time for SO₂ is ≈4 days, so if the average wind velocity is 15 km/h, sulfate is deposited roughly 900 mi downwind of source.

? mi = 4 day
$$\left(\frac{24 \text{ h}}{1 \text{ day}}\right) \left(\frac{15 \text{ km}}{1 \text{ h}}\right) \left(\frac{1 \text{ mi}}{1.609 \text{ km}}\right) = 895 \text{ mi} \approx 900 \text{ mi}$$

Contribution of U.S. Emissions to Total S Deposition in Canada



SO₂ Emissions and Wet Deposition

If 25% of the SO₂ emissions (≈2.5 g(S)/m²/y) are deposited as H₂SO₄, then wet deposition downwind should average ≈0.6 g(S)/m²/y as H₂SO₄. How many kilograms of sulfate. SO₄²⁻, from the sulfuric acid are deposited per hectare per year?

$$\frac{2 \text{ kg SO}_4^{2-}}{\text{ha} \cdot \text{y}} = \frac{0.6 \text{ g S}}{\text{m}^2 \text{ y}} \left(\frac{1 \text{ mol S}}{32 \text{ g S}} \right) \left(\frac{1 \text{ mol SO}_4^2}{1 \text{ mol S}} \right) \left(\frac{96 \text{ g SO}_4^{2-}}{1 \text{ mol SO}_4^{2-}} \right) \left(\frac{10^4 \text{ m}^2}{1 \text{ ha}} \right) \left(\frac{1 \text{ kg}}{10^3 \text{ g}} \right)$$
$$= 18 \text{ kg SO}_4^{2-}/\text{ha} \cdot \text{y}$$

Wet Sulfate Deposition: 1994



National Atmospheric Deposition Program/National Trends Network http://nadp.sws.uiuc.edu

Wet Sulfate Deposition and pH

 H₂SO₄ is a "strong" acid with respect to the loss of the first H⁺. But HSO₄⁻ is a weak acid.

> $H_2SO_4 \rightarrow H^+ + HSO_4^ HSO_4^- \rightleftharpoons H^+ + SO_4^{2-}$

Every mole of H_2SO_4 produces about 1.01 moles of H⁺.

 If the [H⁺] and pH of precipitation are primarily determined by the 0.6 g(S)/m²/y deposited as H₂SO₄ in the ≈1 m/y of precipitation, what is the [H⁺] and pH of the rain?

$$\frac{2 \text{ mol } H^{+}}{L} = \frac{0.6 \text{ g S}}{\text{m}^{2} \text{ y}} \left(\frac{1 \text{ mol } \text{S}}{32 \text{ g S}} \right) \left(\frac{1.01 \text{ mol } H^{+}}{1 \text{ mol } \text{S}} \right) \left(\frac{1 \text{ y}}{1 \text{ m}} \right) \left(\frac{1 \text{ m}^{3}}{10^{3} \text{ L}} \right)$$
$$= 2 \times 10^{-5} \text{ M } H^{+}$$
$$pH = -\log \left[H^{+} \right] = -\log 2 \times 10^{-5} = 4.7$$

A Simple Calculation

- There is a lot more going on here...details of emissions, transport, conversion, and deposition; deposition of other acids (e.g., HNO₃); acidneutralizing reactions with gases (e.g., with NH₃) and aerosols (CaCO₃) ...but this gives a feeling for the relationship between coal burning and the pH of rainfall
- Note:
 - If deposition rate is 2x higher, pH is 0.3 lower (pH = 4.4)
 - If deposition rate 2x lower, pH is 0.3 higher (pH=5.0)

Average pH of Rainfall: 1999

Hydrogen ion concentration as pH from measurements made at the field laboratories, 1999



National Atmospheric Deposition Program/National Trends Network http://nadp.sws.uiuc.edu

Sources of Nitrogen Oxides (NOx)

Anthropogenic 30 Mt(N)/y global, 7 Mt(N)/y US

- Fossil-fuels (21 Mt(N)/y; 70% global, 90+% US)
 - vehicles (45% US)
 - power plants (35% US)
- Biomass burning (2-12 Mt(N)/y)
- Agricultural soils (0-4 Mt(N)/y)

Natural

- Lightning (2-12 Mt(N)/y)
- Soils (3-8 Mt(N)/y)

NO_x is NO and NO_2

- Combustion and other high-temp processes in air (O₂ + N₂) produce NO and NO₂; together they are referred to as "nitrogen oxides" or "NOx"
- 95% of NOx from combustion sources is emitted in the form of NO, but this is quickly oxidized in the atmosphere to NO₂, and so emissions of NOx can be reported as t(NO₂).
- Using t(N) avoids the ambiguity.

 $1 t(N) \times (46/14) = 3.3 t(NO_2)$



Nitric Acid Formation

- Nitric acid formation in gas phase $NO_2 + OH \rightarrow HNO_3$
- HNO_3 formation in water solution $3 NO_2 + H_2O \rightarrow 2 HNO_3 + NO$
- Nitrogen monoxide will oxidize to form nitrogen dioxide that again reacts with water, ultimately forming nitric acid:

 $4 \text{ NO} + 3 \text{ O}_2 + 2 \text{ H}_2\text{O} \rightarrow 4 \text{ HNO}_3$

• Nighttime

 $NO_2 + O_3 \rightarrow NO_3 + O_2$ $NO_3 + NO_2 + M \rightarrow N_2O_5 + M^*$ $N_2O_5 + H_2O \rightarrow 2HNO_3$

Wet Nitrate Deposition: 1994



pH of Acid Rain

- HNO₃ is a "strong" acid: HNO₃ → H⁺ + NO₃⁻
 Every mole of HNO₃ produces a mole of H⁺
- If 0.23 g_N/m²/y as HNO₃ are deposited in ≈1 m/y of precipitation, what is the [H+] and pH of the rain?

 $\frac{? \text{ mol } H^{+}}{L} = \frac{0.23 \text{ g(N)}}{\text{m}^{2} \cdot y} \left(\frac{1 \text{ mol } N}{14.0067 \text{ g(N)}} \right) \left(\frac{1 \text{ mol } H^{+}}{1 \text{ mol } N} \right) \left(\frac{1 y}{1 \text{ mol } N} \right) \left(\frac{1 \text{ m}^{3}}{1 \text{ mol } N} \right)$ $= 1.6 \times 10^{-5} \text{ M } \text{H}^{+}$

 $pH = -log [H^+] = -log (1.6 \times 10^{-5} M) = 4.8$

• If we also have $[H^+] = 2 \times 10^{-5}$ M from sulfate, then pH = -log $(2 \times 10^{-5} + 1.6 \times 10^{-5}) = -log (3.6 \times 10^{-5}) = 4.4$

US NO₂ Emissions, 1940-2002



Transport, Conversion, Deposition



Other Natural Effects on pH of Rain

Natural SO₂ Emissions: $SO_2 + H_2O \leftrightarrow H_2SO_3$ $H_2SO_3 \rightleftharpoons H^+ + HSO_3^ HSO_3^- \rightleftharpoons H^+ + SO_3^{2-}$

bring pH down 0.1 unit (from 5.6 to 5.5)

Natural NH₃ Emissions: NH₃ + H₂O \rightleftharpoons NH₄⁺ + OH⁻

• brings pH up 0.1 unit (from 5.5 to 5.6)

After the Acid Is Deposited

• Evaporation and concentration of wet-deposited acids (on leaf and rock surfaces, in acid snows)



Acid Shock

- Episodic acidification refers to brief periods during which pH levels decrease due to runoff from melting snow.
- Occurs in mid-latitudes where acid snow accumulates in winter, concentrate through evaporation, and melts rapidly in spring.
- Acids are released over a short period of time at concentrations 5 to 10 times more acidic than rainfall.

Acid Neutralization

Reaction with natural bases

 $CaCO_{3} + 2H^{+} \rightleftharpoons Ca^{2+} + CO_{2} + H_{2}O$ $AI(OH)_{3} + 3H^{+} \rightleftharpoons AI^{3+} + 3H_{2}O$

• Nutrient leaching:

 $K-[soil] + H^+ \rightleftharpoons H-[soil] + K^+$

Carbonate Reactions

• Weathering reactions with carbonate minerals (calcite, dolomite, various limestones, etc.):

 $CaCO_{3} + CO_{2} + H_{2}O \rightleftharpoons Ca^{2+} + 2HCO_{3}^{-}$ $CaSiO_{3} + 2CO_{2} + H_{2}O$ $\rightleftharpoons Ca^{2+} + SiO_{2} + 2HCO_{3}^{-}$

• When acid is added to water, equilibria shift and bicarbonate neutralizes some of the acid:

 $HCO_3^- + H^+ \rightleftharpoons H_2O + CO_2$

 If soils and sediments are rich in carbonates, weathering will supply additional HCO₃⁻.

Alkalinity

- Alkalinity is a measure of the "buffering" or "acid-neutralizing capacity" (ANC) of a solution:
 ANC = [HCO₃⁻] + 2[CO₃²⁻] + [OH⁻] – [H⁺]
- Water saturated with CaCO₃:

ANC = 970×10^{-6} Eq/L = 970 µeq/L

- For bases, an equivalent (Eq) is the amount of a substance which will react with one mole of hydronium ions (H₃O⁺).
- 1 mol $CO_3^{2-} = 2 Eq CO_3^{2-}$
- ANC depends on watershed soil characteristics.

Buffer Solutions

- A buffer is an aqueous solution a weak acid (H₂CO₃) and its conjugate base (HCO₃⁻) or a weak base (CO₃²⁻) and its conjugate acid (HCO₃⁻).
 - Conjugate base = acid less one H⁺
 - Conjugate base = base plus one H⁺

```
H_2CO_3(aq) \rightleftharpoons H^+(aq) + HCO_3^-(aq)
```

$$K_{a1} = \frac{[H^+][HCO_3^-]}{[H_2CO_3]}$$
$$[H^+] = K_{a1} \frac{[H_2CO_3]}{[HCO_3^-]}$$
$$[H^+] = K_a \frac{[weak acid]}{[conjugate base]}$$

 As long as there are significant concentrations of the weak acid (H₂CO₃) and the conjugate base (HCO₃⁻), changing one or the other does not change the ratio of acid to base much so the [H⁺] and pH stay relatively constant.
Simplistic Calculation to Illustrate Buffering

- pH pristine natural water is about 5.6.
- Calculate the pH of a solution where enough strong acid is added to water to yield a 0.1 M H⁺ solution.
 pH = -log(10⁻¹) = 1
- Calculate the pH of a solution of 1 M H_2CO_3 and 1 M HCO_3^- .

$$K_{a1} = \frac{[H^+][HCO_3^-]}{[H_2CO_3]} = 4.5 \times 10^{-7}$$
$$[H^+] = 4.5 \times 10^{-7} \frac{[H_2CO_3]}{[HCO_3^-]} = 4.5 \times 10^{-7} \frac{1 \text{ M H}_2CO_3}{1 \text{ M H}CO_3^-}$$

 $[H^+] = -\log 4.5 \times 10^{-7} = 6.35$

Simplistic Calculation to Illustrate Buffering

Calculate the pH of a solution of 1 M H₂CO₃ and 1 M HCO₃⁻ when enough strong acid is added to the solution to make the solution 0.1 M strong acid.

$$HCO_{3}^{-1}(aq) + H^{+}(aq) \rightarrow H_{2}CO_{3}(aq)$$

 $[HCO_{3}^{-1}] = 1 \text{ M} - 0.1 \text{ M} = 0.9 \text{ M} \text{ HCO}_{3}^{-1}$
 $[H_{2}CO_{3}] = 1 \text{ M} + 0.1 \text{ M} = 1.1 \text{ M} \text{ H}_{2}CO_{3}$

$$\begin{split} \mathsf{K}_{a1} &= \frac{[\mathsf{H}^+][\mathsf{HCO}_3^-]}{[\mathsf{H}_2\mathsf{CO}_3]} = 4.5 \times 10^{-7} \\ [\mathsf{H}^+] &= 4.5 \times 10^{-7} \ \frac{[\mathsf{H}_2\mathsf{CO}_3]}{[\mathsf{HCO}_3^-]} = 4.5 \times 10^{-7} \ \frac{1.1 \ \mathsf{M} \ \mathsf{H}_2\mathsf{CO}_3}{0.9 \ \mathsf{M} \ \mathsf{HCO}_3^-} \end{split}$$

 $[H^+] = -\log 5.5 \times 10^{-7} = 6.26$

Buffer Solution – Base and Conjugate Acid

 A lake might contain a buffer solution of the base carbonate from the solution of calcium carbonate in limestone and hydrogen carbonate in runoff from weathering of the limestone.

 $CaCO_{3}(s) \rightleftharpoons Ca^{2+}(aq) + CO_{3}^{2-}(aq)$ $CaCO_{3} + CO_{2} + H_{2}O \rightleftharpoons Ca^{2+} + 2HCO_{3}^{-}$

• Many life forms thrive only in a relatively small pH range so buffer solutions are important for maintaining a near constant pH.

Alkalinity

- Surface water with ANC ≈200 µeq/L is normal; healthy lakes have ANC > 100 µeq/L; water with ANC < 50 µeq/L is highly sensitive to acidification.
- Acidified waters have ANC < 0 μeq/L.
- As acidity [H⁺] is added to a lake, ANC is consumed. Because of the buffering, the pH decreases slowly at first, but as the basic components of the buffer react with the acid, the water loses its buffering ability, and the pH drops more rapidly.

pH vs. ANC



Acidification of a Lake

Assume small lake with a volume of 10¹⁰ L, a watershed area of 1 km², precipitation of 1 m/y with pH = 4 or [H+] = 10⁻⁴ M. If soils (e.g., granite) do not neutralize the acid, flow of acid in mEq H⁺ per liter per year into the lake is:

$$\frac{2 \,\mu Eq \,H^{+}}{L \cdot y} = \frac{1 \,km^{2}}{10^{10} \,L} \left(\frac{10^{3} \,m}{1 \,km}\right)^{2} \left(\frac{1 \,m}{1 \,y}\right) \left(\frac{10^{3} \,L}{1 \,m^{3}}\right) \left(\frac{10^{-4} \,mol \,H^{+}}{1 \,L}\right) \left(\frac{1 \,Eq \,H^{+}}{1 \,mol \,H^{+}}\right) \left(\frac{10^{6} \,\mu Eq}{1 \,Eq}\right)$$
$$\approx 10 \,\mu Eq \,H^{+}/L \cdot y$$

ANC would decrease by 10 μeq/L per year.

Acidification of a Lake



Liming the Lake

- After the lake is acidified, we can neutralize the acid by adding lime (CaCO₃) CaCO₃ + CO₂ + H₂O \Rightarrow Ca²⁺ + 2HCO₃⁻
- To raise ANC by 200 $\mu eq/L,$ we would add:

?
$$t \operatorname{CaCO}_{3} = 10^{10} \mathcal{L} \left(\frac{200 \times 10^{-6} \operatorname{Eq}}{1 \, \mathcal{L}} \right) \left(\frac{1 \operatorname{mol} \operatorname{CaCO}_{3}}{2 \operatorname{Eq}} \right) \left(\frac{100.087 \, g \operatorname{CaCO}_{3}}{1 \operatorname{mol} \operatorname{CaCO}_{3}} \right) \left(\frac{1 \, t}{10^{6} \, g} \right)$$

= 100 t CaCO₃

 If the acid flow into the lake continues, this would have to be replenished on a time scale comparable to the residence time of water in the lake (in this case, ≈10 y).

Sensitive Areas



Source: NAPAP. 1991, 1990 Integrated Assessment Report.

Sensitive Areas



Critical Load

- The maximum amount of acid deposition that a region can receive without damage to its ecosystems is known as its "critical load".
- CL depends on the ANC of water, rocks, and soils.



$$\frac{24 \text{ kg}(\text{SO}_4)}{\text{L}} = \frac{24 \text{ kg}(\text{SO}_4)}{\text{ha} \cdot \text{y}} \left(\frac{10^3 \text{ g}}{1 \text{ kg}}\right) \left(\frac{1 \text{ mol}(\text{SO}_4)}{96.0636 \text{ g}(\text{SO}_4)}\right) \left(\frac{2 \text{ mol} \text{ H}^+}{1 \text{ mol}(\text{SO}_4)}\right) \left(\frac{1 \text{ y}}{1 \text{ m}}\right) \left(\frac{1 \text{ ha}}{10^4 \text{ m}^2}\right) \left(\frac{1 \text{ m}^3}{10^3 \text{ L}}\right)$$
$$= 5.0 \times 10^{-5} \text{ M H}^+$$

$$pH = -log [H^+] = -log (5.0 \times 10^{-5} M) = 4.3$$

Nutrient Leaching

- Soil chemistry is governed by negative charge of clay and humus particles, which gives them the ability to retain and exchange cations
- Each ion can displace any ion to its right:

 $H^+ > Ca^{2+} > Mg^{2+} > NH_4^+ > K^+ > Na^+$

• Because H⁺ binds more strongly with the soil particles than nutrients, H⁺ displaces nutrients:

 $K-[soil] + H^+ \leftrightarrow H-[soil] + K^+$

K⁺ ion is then dissolved in water; unless taken up by plants, it will be leached out

Nutrients in Runoff: Czechoslovakia, 1976-82								
	Concentratio Runoff (k							
	Undamaged	Damaged	Ratio					
K Potassium	1.9	6.8	3.6					
Mg Magnesium	3.8	26	6.8					
Ca Calcium	9.9	74	7.5					
S Sulfate	9.0	96	10.7					
N Nitrate	0.6	12	20					
AI Aluminum	0.1	3.2	32					

Aluminum Toxicity

• Aluminum hydroxide solid is a base that can react with acid. For example:

 $AI(OH)_3(s) + 3H^+(aq) \leftrightarrow AI^{3+}(aq) + 3H_2O(I)$

- The Al³⁺ ion is then dissolved in the runoff, which washes into streams and lakes.
- Typical AI concentrations in acidified lakes are 400 to 700 μ g/L, increasing with decreasing pH.
- Both low pH and increased aluminum levels are directly toxic to fish.
- In addition, low pH and increased aluminum levels cause chronic stress that may not kill individual fish, but leads to lower body weight and smaller size and makes fish less able to compete for food and habitat.



Impacts

- Freshwater ecosystems
- Forests
- Agriculture
- Human health
- Buildings and property
- Visibility

Freshwater Ecosystems

- Areas with combination of high deposition and low ANC are the first effect observed from acid.
- Ecosystem collapsed in ≈15,000 lakes in Scandinavia; serious damage in ≈14,000 in Canada (Ontario), and ≈14,000 in US (Adirondacks)
- Acidity reduces productivity of aquatic plants, microbial activity (decomposition, nutrient cycling)
- At moderate pH (5-6), harm to fish primarily through disruption of food chain and reproductive cycle, not direct damage.
- At low pH, toxic metals (esp. AI) may play a role.

Acidification of lakes in southern Norway

Percent of surveyed lakes with given pH



Acid Tolerance of Freshwater Animals

	⊨H 6.5	⊨H 6 .0	eH 5.5	⊨H 5 .0	PH 4.5	⊨H 4 .0
трол						
BASS						
PERCH						
FROGS						
SALAMANDERS						
CLAMS						
CRAYFISH						
SNAILS						
MAYFLY						

Loss of Fish Stocks in Norway



pH and Number of Species: Adirondacks



ANC and Fish Species: Adirondacks



Forests and Acid Rain



Forests and Plants

- Forest damage first noticed in Germany in 1970s
 - In Europe, 15% of standing forest damaged
- Acid damage appears to be due primarily to dry deposition and acid fogs
- Also: leach nutrients, suppress microbes (nutrient cycling), mobilize toxic metals (inhibit water, nutrient uptake by fine roots)
- In U.S., Scandinavia, no clear correlation between wet deposition and damage
- Other explanations: disease, pests, ozone, UVB, climate change (more vulnerable due to acid?)



Human Health

- Acids
 - Consumption of fish with elevated concentrations of toxic metals (AI, Pb, Cd, Hg, Cu, Zn)
 - Corrosion of pipes resulting in excess levels of Cu, Zn, Pb
 - In Adirondacks, Pb levels up to 100 × standard
- Acid precursors
 - SO₂ and NO_x are toxic;
 - NO_x react to form ozone
 - Respiratory and cardiac diseases

Buildings and Property

• Carbonate stones (marble, limestone, etc.), cement, mortar are dissolved by acids:

 $CaCO_3(s) + H_2SO_4(aq)$

 \rightarrow CaSO₄(aq) + CO₂(g) + H₂O(l)

- Acids etch glass; damage roofing and other building materials; damage plastics and paint (esp. automotive paint)
- Damage to U.S. buildings: ≈\$5 billion/y

Capitol Building, Ottawa





Marble surfaces develop a rough "sugary" texture because calcite grains are loosened as the edges dissolve

Effects on Metals



Mitigation

- Tall Stacks (export pollution)
- Sulfur
 - Switch from coal to natural gas (0.001% S)
 - Switch to low-sulfur coal
 - S content ranges from 0.3 to 6%; only 8% of US coal would meet 1975 NSPS (<0.7% S)
 - Scrub coal (solvent refining removes up to 75% S)
 - Flue-gas desulfurization (FGD)
 - mix exhaust gas with lime (CaO); SO₂ reacts for form gypsum (CaSO₄)

SO₂ Emissions Reduction

- Due largely to the US EPA's Acid Rain Program, the U.S. had a 33% decrease in SO₂ emissions between 1983 and 2002.
- Partly from flue-gas desulfurization.



Wet SO₂ Scrubber



Flue-Gas Desulfurization

- A technology that enables SO₂ to be removed in power plants burning sulfur-containing coal or oil.
- Wet scrubbing uses a CaCO₃ (limestone) slurry and produces CaSO₃ (calcium sulfite):

 $CaCO_3(s) + SO_2(g) \rightarrow CaSO_3(s) + CO_2(g)$

 When wet scrubbing with a Ca(OH)₂ (lime) slurry, the reaction also produces CaSO₃ is:

 $Ca(OH)_2(s) + SO_2(g) \rightarrow CaSO_3(s) + H_2O(I)$

 When wet scrubbing with a Mg(OH)₂, the reaction produces MgSO₃:

 $Mg(OH)_2(s) + SO_2(g) \rightarrow MgSO_3(s) + H_2O(I)$

 To partially offset the cost of the FGD installation, in some designs, the CaSO₃ is further oxidized to produce marketable CaSO₄ · 2H₂O (gypsum).

 $CaSO_3(s) + \frac{1}{2}O_2(g) \rightarrow CaSO_4(s)$

Mitigation - Coal Gasification

- A gasifier converts hydrocarbon feedstock into gaseous components by applying heat under pressure in the presence of steam.
- Only a relatively small portion of the fuel burns completely. This "partial oxidation" process provides the heat.
- Produce "syngas", primarily hydrogen and carbon monoxide

 $C + H_2O \rightarrow CO + H_2$ $CO + H_2O \rightarrow CO_2 + H_2$

- Sulfur converted to hydrogen sulfide, H₂S, and carbonyl sulfide, OCS, which can be converted into elemental sulfur or sulfuric acid, both valuable byproducts.
- 99+% of S removed
Coal Gasification



Automobile Catalytic Converters

Catalytic converters: up to 95% reduction in NO_x 2NO \rightarrow N₂ + O₂ 2NO₂ \rightarrow N₂ + 2O₂



Standards

 National Ambient Air Quality Standards (NAAQS) – part of Clean Air Act (CAA)

http://www.epa.gov/air/criteria.html

http://www.epa.gov/air/caa/

- Two way to achieve standards:
 - Command-and-control
 - Emission and technology standards
 - Market-based mechanisms
 - Taxes and tradable permits
 - Cap-and-trade with free permits: guaranteed reduction, property rights to existing polluters

Clean Air Act Amendments of 1990

- Title I: NAAQS and schedule for attainment
- Title II: Emission standards for mobile sources

Title IV: Acid Deposition Control

- 10 Mt(SO₂) reduction below 1980 level by 2007; coal-fired utilities from $17.5 \rightarrow 8.5$ Mt
 - Phase I: 110 high-emitting plants (10.9 \rightarrow 8.7 Mt)
 - Phase II: all plants > 25 MW, 8.5 Mt cap
- 2 Mt(NO_x) reduction in 2 phases, based on emission rates, achieved w/low-NO_x burners
 - cost = \$230/t v. \$7000/t for cars

Tradable Permit System for SO₂

- Beginning in 1995, emissions ≤ allowances
- Allocation:
 - existing plants: 2.5 lb/MBTU (Phase I), 1.2 lb/MTU (Phase II), based on 1985-87 average emissions
 - new plants must purchase permits
 - 2.8% withheld for auction by EPA
 - direct sale @ \$1500/t
- Unused allowances can be traded, banked
- Voluntary opt-in for other sources
- Continuous emission monitoring; allowance tracking
- Penalty = \$2000/t <u>plus</u> allowances
- Ignores location of sources

SO₂ Emissions from Acid Rain Program Sources, 1980–2009



http://www.epa.gov/airmarkets/progress/ARP09_1.html

SO₂ Emissions and the Allowance Bank, 1995–2009



http://www.epa.gov/airmarkets/progress/ARP09_1.html

NO_x Emission Trends for All Acid Rain Program Units, 1990–2009



http://www.epa.gov/airmarkets/progress/ARP09_1.html

1995 SO₂ Emissions



2004 SO₂ Emissions



Overall, Title IV Considered Success

- Flexibility: low-S coal, scrubbing, fuel switching, efficiency, conservation;
- Costs << command-and-control
- Benefits >> costs; mostly health-related
- Observed reduction in S deposition, increase in pH, slow recovery of some NE lakes

But additional reductions necessary:

- No recovery, continued acidification of Adirondack lakes, southern Blue Ridge streams
- NO_x more important than originally thought